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1 **Human appropriation of land for food: the role of diet**

2 **Abstract**

3 Human appropriation of land for food production has fundamentally altered the Earth system, with
4 impacts on water, soil, air quality, and the climate system. Changes in population, dietary preferences,
5 technology and crop productivity have all played important roles in shaping today's land use. In this
6 paper, we explore how past and present developments in diets impact on global agricultural land use.
7 We introduce an index for the Human Appropriation of Land for Food (HALF), and use it to isolate the
8 effects of diets on agricultural land areas, including the potential consequences of shifts in consumer
9 food preferences. We find that if the global population adopted consumption patterns equivalent to
10 particular current national per capita rates, agricultural land use area requirements could vary over a
11 14-fold range. Within these variations, the types of food commodities consumed are more important
12 than the quantity of per-capita consumption in determining the agricultural land requirement, largely
13 due to the impact of animal products and in particular ruminant species. Exploration of the average
14 diets in the USA and India (which lie towards but not at global consumption extremes) provides a
15 framework for understanding land use impacts arising from different food consumption habits.
16 Hypothetically, if the world were to adopt the average Indian diet, 55% less agricultural land would be
17 needed to satisfy demand, while global consumption of the average USA diet would necessitate 178%
18 more land. Waste and over-eating are also shown to be important. The area associated with food
19 waste, including over-consumption, given global adoption of the consumption patterns of the average
20 person in the USA, was found to be twice that required for all food production given an average Indian
21 per capita consumption. Therefore, measures to influence future diets and reduce food waste could
22 substantially contribute towards global food security, as well as providing climate change mitigation
23 options.

24

25

26

1. Introduction

Human appropriation of global net primary production (NPP) of vegetation is increasing, and has doubled since 1910 (Krausmann et al., 2013). This is due to rising populations, as well as changes in diets. Diet is linked with wealth (Tilman et al., 2011), urbanisation (Huang and Bouis, 2001; Seto and Ramankutty, 2016; Wu and Wu, 1997), and globalising food commodity markets (Pingali, 2007; Popkin, 2006; Yu et al., 2013). These changes, including rising incomes, have seen a concomitant increase in food consumption and shift towards higher rates of consumption of commodities that are more land-intensive to supply; in particular meat and milk (Godfray et al., 2010; Tilman and Clark, 2014; Weinzettel et al., 2013).

Shifts in diets have become an increasingly important driver for land use change over time (Alexander et al., 2015; Kastner et al., 2012), a process that is likely to continue even as the rate of population growth slows (van Vuuren and Carter, 2014). Although increases in yields and production efficiencies have offset additional demand for food commodities, agricultural land areas have been expanding (FAOSTAT, 2015a). Environmental impacts can occur either through the expansion of agricultural production and consequent loss of a previous land cover, or through the intensification of production, e.g. eutrophication or biodiversity loss (Smith et al., 2013). Land use and the environmental impacts associated with agricultural production are also increasingly displaced from the country of consumption, through international trade of food commodities (Erb et al., 2009; Weinzettel et al., 2013; Yu et al., 2013). Agriculture accounts for around a third of global anthropogenic greenhouse gas (GHG) emissions, and land-use change alone presently accounts for 10% of anthropogenic CO₂ emissions (Le Quéré et al., 2015). As well as causing environmental issues, dietary transitions have contributed to rising global rates of obesity and increases in associated diseases, e.g. diabetes and heart disease (Hu, 2011; Tilman and Clark, 2014).

Animal products contribute disproportionately low amounts of energy and protein to human diets (respectively 18 and 39 % globally in 2011), relative to their land-use footprint (pasture accounts for approximately 68% of agricultural land, plus around one third of cropland is used for the production of animal feeds (Alexander et al., 2015; FAO, 2006)). However, grassland is a broad category that covers a diverse range of intensities, from intensively managed pasture to extensively used savannahs with little or no inputs of fertiliser or other management, meaning that direct comparisons between different land use areas are difficult. Nonetheless, the expansion of pasture (62% of the expansion in agricultural area from 1961 to 2011 (FAOSTAT, 2015a)), as well as the increasing use of crops for feed, demonstrates the critical importance of animal products as a driver of land use change. Animal products also play a role in water consumption (Jalava et al., 2014), and agricultural GHG emissions not associated with land use change (Tilman and Clark, 2014). The impacts from food production, both of animal products and crops, are exacerbated by losses or inefficiencies that exist at each stage in the production system, from harvesting, through transport and storage, to processing and finally at the consumer (Gustavsson et al., 2011; Parfitt et al., 2010).

Future food requirements could be met through a combination of increasing production and reducing demand. However, substantial attention has been given to supply-side responses, including expanding land in agricultural use and increasing food yields, especially crops (e.g. closing the 'yield gap' or 'sustainable intensification') (Foley et al., 2011; Kastner et al., 2014; Mueller et al., 2012; West et al., 2014); or the potential benefits and trade-offs associated with increasing livestock intensities (Davis et al., 2015; Herrero et al., 2016). Such analyses tend to consider dietary change as an exogenous wealth-based factor, and anticipate continuations of current dietary trends (Engström et al., 2016; Schmitz et al., 2014). However, diets and the food preferences that shape them do not necessarily follow fixed

1 trends. Instead, they alter over time influenced by technology, policies and changes in social norms,
2 e.g. (Hollands et al., 2015). Modelling work has been done to project the impact of alternative
3 assumptions regarding future diets (Bajželj et al., 2014; Haberl et al., 2011; Popp et al., 2010; Stehfest
4 et al., 2009), and the ability of the agricultural system to supply the global population with a diet
5 containing adequate calories has also been considered (Cassidy et al., 2013; Davis et al., 2014). Further
6 studies in this area have taken a life-cycle analysis (LCA) approach that typically consider either GHG
7 emissions, energy or water requirements for individual commodities (Carlsson-Kanyama and González,
8 2009; González et al., 2011; Marlow et al., 2009; Pelletier et al., 2011). However, few studies have
9 quantified the impact of variations in existing diets. Erb *et al.* (2009) considered the impact of current
10 variations in food consumption patterns on agricultural land use, by quantifying trade in the embodied
11 human appropriation of biomass net primary production. But, despite the potential significance of
12 consumer behaviours on land use, no attempt appears to have been made to quantify the land use
13 impacts of existing diets, dissociated from the complicating effect of domestic production and
14 international trade.

15
16 Here, we address this gap by proposing a new index and using it to quantify the land use requirements
17 of diets by country and over time (from 1961 to 2011). The Human Appropriation of Land for Food
18 (HALF) index expresses the land area required for the global population to consume a particular diet, as
19 a percentage of the world land surface. HALF therefore provides a relative measure of the scale of the
20 impacts of alternative diets on land use. Diet here is assumed to include the quantities of commodities
21 lost and wasted after reaching the consumer. The index is calculated from global average production
22 intensities and yields from a baseline year, primarily 2011. HALF is accordingly not predictive, as
23 adaptive responses in production systems that may result from changes in demand are excluded.
24 Rather, the HALF index is a metric that characterises the land use impact of alternative scenarios of
25 dietary patterns. The results can be interpreted in terms of both methods and areas of production,
26 with a given increase in the HALF index implying the same increase in agricultural areas, an equivalent
27 increase in productive efficiency, or some combination of the two.

28

29 **2. Method**

30 FAO country-level panel data for crop areas, production quantities, commodity uses and nutrient values
31 were used to construct the HALF index (FAOSTAT, 2015a, 2015b, 2015c, 2015d, 2015e, 2015f). Global
32 average production values and efficiencies for primary crops, processed commodities and livestock
33 products were used to calculate the agricultural areas needed to meet per capita consumption for each
34 country. The index is expressed as the percentage of the world's land surface required for the global
35 population to adopt each country's diet. All diets are evaluated using the global average production
36 system. Assessments of country average diet do not use production or international trade associated
37 with that country, except as they contribute to the world average. The calculations and assumptions
38 are described in more detail below, with a summary of assumptions available in Table S2.

39

40 *(a) Allocating areas for food commodities*

41 The areas associated with the production of 90 commodities (see Table S3), representing 99.4% of
42 global food consumption by calorific value, were each allocated between three categories of use: food
43 for human consumption, animal feed, and non-food related uses (primarily biofuels and fibre). The
44 commodities comprise 50 primary crops that are directly grown, 32 processed commodities derived
45 from them, and 8 livestock products. The FAO commodity balance data (FAOSTAT, 2015d) identifies
46 the quantities used for food, feed, processing, other non-food related uses (primarily, bioenergy and
47 fibre), seed and waste. To provide an assessment of the embedded areas required for delivering the
48 consumed commodities two adjustments were made. Firstly, for each primary crop, the quantities

1 used as seed and wasted (e.g. in storage and transport) were distributed across the remaining
2 categories of use (i.e. food, feed, processing and non-food). The second adjustment deals with the
3 difference between the total cropland area and the harvested areas (e.g. in 2011, respectively, 1556
4 Mha and 1378 Mha (FAOSTAT, 2015a, 2015c)) due to set-aside, multiple-cropping, and failed or
5 unharvested crops. To account for these differences, the cropland area for each primary crop was
6 adjusted by the ratio of these areas (e.g. in 2011 areas they are increased by a factor of 1.129). After
7 applying both the adjustments, the cropland area for each primary crop was then allocated pro-rata
8 between the categories of use (i.e. food, feed, processing and non-food), by the mass used for each
9 category. This approach removes the areas used to produce commodities for bioenergy, fibre or other
10 non-food uses. Example calculations are given in the SI Methods.

11
12 The areas used to grow the primary crops for processing were further mapped to the commodities
13 output from the processing. Where multiple commodities are produced from a single crop, the areas
14 used to grow the primary crop were allocated on an approximate economic value basis (Table S4). For
15 example, processed oil crop areas were divided equally between the resulting oil (used primarily for
16 food and biofuel), and the seed meals or cakes (used primarily for livestock feed). In 2011, 224.1 Mt of
17 soybeans, which represent the single biggest vegetable oil crop (48% of the total), were processed
18 globally into 41.6 Mt of oil and 174.7 Mt of meal (7.8 Mt is assumed lost during processing). This gives
19 a similar total market value for the oil and meal (45% of value is in the oil and 55% in meal), at 2011
20 market prices of \$1103/t and \$321/t respectively (Index Mundi, 2016), suggesting that an equal division
21 of input area is a reasonable approximation. Alternative allocations would introduce additional biases.
22 For example, calculations on the basis of mass would be biased towards associating the area with the
23 seed meals, while conversely accounting for them as a by-product with no area allocated would
24 implicitly and incorrectly assume they can be freely produced and have no value.

25

26 *(b) Allocating areas for animal feed and pasture*

27 Animal nutrition derives from grassland and feed crops including forage crops. Data are available to
28 quantify the area of pasture and quantities of crops used as feed (FAOSTAT, 2015a, 2015d). However,
29 there are no empirical data to describe directly how these sources of nutrition are divided between
30 livestock species, and hence between commodity types such as meat, milk and eggs. Instead, feed
31 conversion ratios (FCRs), describing the efficiency of converting inputs into edible animal products,
32 were used to estimate animal feed requirements (Table 1). Commonly, FCRs are expressed in terms of
33 dry matter (DM) of feed per animal live weight (LW). To represent the production efficiency of meat
34 consumed by humans, these ratios were adjusted to express feeding requirements per unit edible
35 weight (EW), and also to account for the need to raise sire and dam animals (Smil, 2002).

36

37 The nutritional requirements of monogastric livestock (i.e. poultry and pigs) were assumed to be met
38 solely from feed, while nutrients for ruminant species (e.g. cattle and sheep) come from feed and
39 grazed pasture. Firstly, the produced masses from monogastric animals were multiplied by the feed
40 conversion factors (Table 1) to give estimates of the feed requirements. These feed amounts, and the
41 cropland areas needed to provide them, were allocated to the monogastric livestock products.
42 Secondly, the remaining feed (23% in 2011 using feed dry matter content (INRA et al., 2016)), and
43 associated cropland areas were allocated *pro rata* by the estimated feed requirements across the
44 ruminant products. The same *pro rata* allocation was used to associate the pasture area with products
45 derived from ruminant animals. See SI Methods for a worked example.

46

1 *Table 1. Global average feed conversion ratios and efficiencies for animal products. The feed*
 2 *conversion efficiencies and direct energy for housing are given for reference, and are not used in the*
 3 *analysis.*

Animal product	Feed conversion ratio (kg DM feed/kg EW)	Percentage edible (% EW of LW)	Energy feed conversion efficiency (%)	Protein feed conversion efficiency (%)	Direct energy for housing or processing (MJ / kg EW)	Data source
Poultry	3.3	70	13	19.6	4.5	(Macleod et al., 2013; Smil, 2013)
Pork	6.4	55	8.6	8.5	1.8	(Macleod et al., 2013; Smil, 2013)
Beef	25	40	1.9	3.8	0.08	(Opio et al., 2013; Smil, 2013)
Other meat *	15	55	4.4	6.3	0.09	(Opio et al., 2013; Smil, 2013)
Eggs	2.3	-	19	25	1.3	(Macleod et al., 2013; Smil, 2013)
Whole Milk	0.7	-	24	24	0.22	(Little, 2014; Opio et al., 2013)
Notes: * The 'other meats' category, which forms 6.6% of all meats produced in 2011, is based on sheep and goat meat (65% by mass of 'other meat' in 2011), but includes other sources of meats, e.g. horse, rabbit and camelids.						

4
 5 *(c) Assessing the land use impact of different diets*
 6 The average consumption per capita and per commodity were calculated globally and nationally
 7 (FAOSTAT, 2015b, 2015d). The area required to produce each commodity was determined from the
 8 global production system land use allocations (described above). The area needed to provide all the
 9 commodities for each country's diet if it were adopted by the global population could then be
 10 calculated (FAOSTAT, 2015g). This was expressed as a proportion of total global land area to obtain the
 11 Human Appropriation of Land for Food (HALF) value. HALF values were also calculated to quantify the
 12 land use impacts of changes in country-level diets over time. The values primarily used here were
 13 calculated with variable diet only, and a constant baseline population and production system (2011 was
 14 chosen as the most recent year with available values (FAOSTAT, 2015d)).

15
 16 National land footprints for food, i.e. an estimate of the actual agricultural land area used supply to
 17 each country's food, were also calculated based on domestic production and the land displaced
 18 through international trade. This used the same data as the HALF calculation, and accounted for
 19 imports and exports following the approach of previous studies (Alexander et al., 2015; Jalava et al.,
 20 2014). For each commodity, net exports were included using the domestic production yields, and net
 21 imports using the global mean yields of net exports (weighed by net export quantities). The country
 22 footprints were expressed as an area per capita using country populations (FAOSTAT, 2015g).
 23 Expressing as a fraction of global land area required for the global population, to match HALF values,
 24 could not be justified as the land footprints are country specific (e.g. in climate and soil).
 25

1 *(d) Decomposing dietary changes into quantities consumed and commodity profiles*

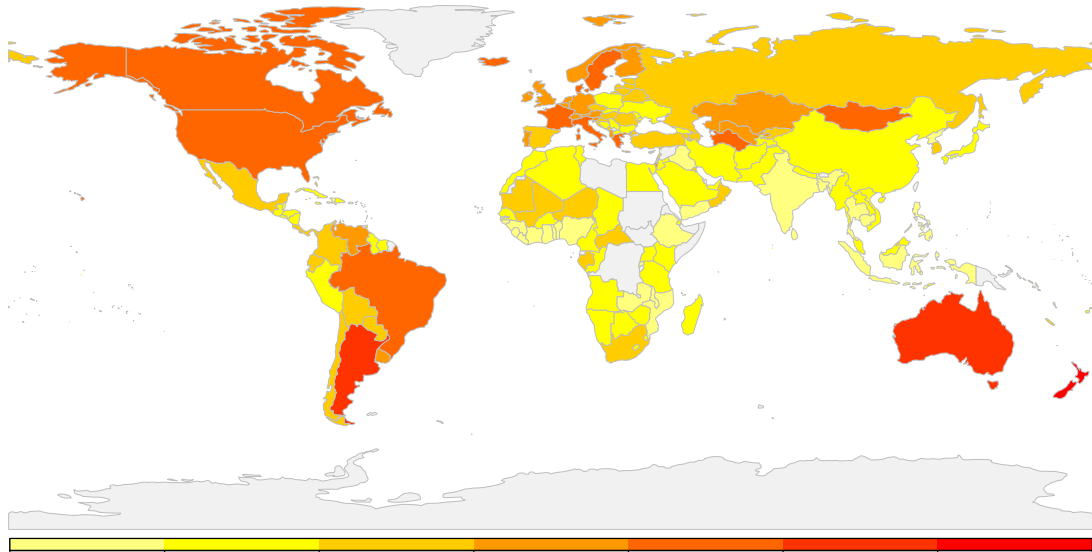
2 The impacts of potential shifts in diets from the 2011 global baseline to that of a particular country was
3 decomposed into two parts. The first part represents a shift in the total quantity of nutrients consumed
4 while holding the proportional contribution of each commodity constant. The second part represents a
5 shift in the ratio or profile of commodities consumed, while holding the total nutrient level constant.
6 These two parts were expressed both in protein and energy terms, with nutritional values by mass for
7 each commodity derived from global FAO food supply data (FAOSTAT, 2015e, 2015f). For example, the
8 average energy consumed per capita globally is 11.9 MJ/person/day, while in the USA the average is
9 16.6 MJ/person/day, i.e. 40% more. Therefore, if the current global profile commodities remained
10 unchanged, but the energy consumed increased to that of the USA, 40% more land would be required
11 for production, in the absence of production intensification. This is reflected in a 40% increase in HALF.
12 However, consumption in the USA also differs in the relative profile of the different commodities
13 consumed. These differences also have an effect on the land required, evaluated without the influence
14 of the quantity differences in the 'profile' type.
15

16 **3. Results**

17 *(a) Global and country-level HALF*

18 The total agricultural area used for human food production was 4484 Mha in 2011, of which 871 Mha
19 was used for cropland for human consumption, and 3700 Mha for animal products (497 Mha of
20 cropland for feed and 3203 Mha of pasture). The remaining cropland was used for biofuels (140 Mha),
21 fibre (33 Mha), feed for non-food uses of animal products (9 Mha), and net variations in stock levels (7
22 Mha). Expressed as a percentage of the global land surface (13,009 Mha (FAOSTAT, 2015a)) the Human
23 Appropriation of Land for Food (HALF) index is 35.1, or an average area per person of 0.65 ha.
24 Expressing HALF as a percentage of global land surface includes land that is unlikely to be suitable for
25 agriculture, e.g. ice-covered or desert areas. However, the use of an estimate of suitable land suffers
26 from difficulty in definition and measurement, and also would vary with climate change. Consequently,
27 the clarity of comparing to the global land surface was preferred.
28

29 There are large differences in HALF values between country-level average diets. For example, the
30 global adoption of the diet in the USA would require over 6 times the agricultural area that adoption of
31 the diet in India, with a HALF index of 97.7 compared to India's 15.8. Figure 1 shows the HALF index at
32 2011 for the average diets of 170 countries for which sufficient data were available (Table S5). The
33 highest HALF values are for diets in New Zealand, Argentina and Australia at 135.8, 114.9 and 112.2
34 respectively, due to the high levels of animal products – particularly beef - consumed. At the other
35 extreme are Mozambique, Liberia, Bangladesh and Sri Lanka all with a HALF index below 11.5, i.e. less
36 than a third of the global average.
37

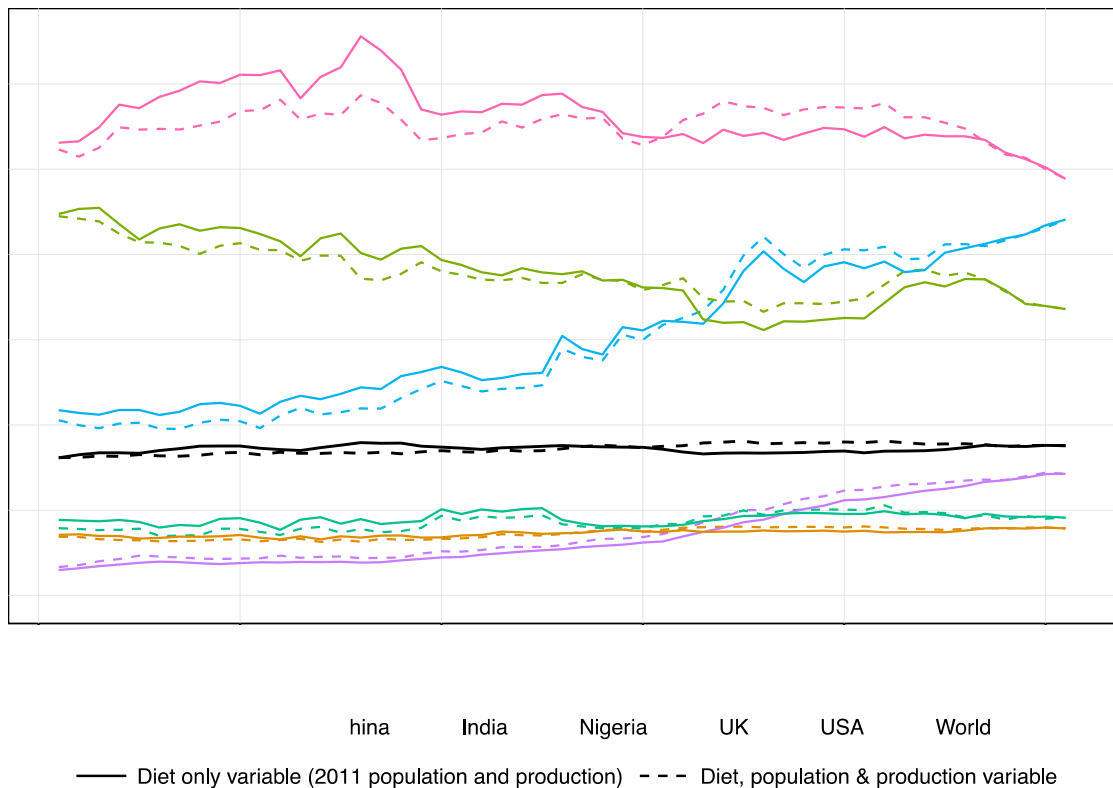


1
 2 *Figure 1. Map of HALF index for average country-level diets in 2011. Countries where the index could*
 3 *not be calculated due to no commodity consumption data being available (FAOSTAT, 2015d), e.g. Libya,*
 4 *Somalia and Greenland, are shown in light grey.*

5
 6 The HALF results use global mean production efficiencies, and so no specific account is taken of
 7 domestic (national) production except as it contributes to the world average. The national food
 8 footprints (Figure S1) include aspects of diet and production within them, whereas HALF (Figure 1) only
 9 includes variations in diet. The distribution of these national footprints differ from the distribution of
 10 HALF values as a result (e.g. Mongolia has a per capita footprint 3 times greater than any other country
 11 (39 ha/person), due to the use of extensive grazing). Many developed countries have a lower land use
 12 footprint than implied by the HALF index, due to the high agricultural yields in these countries. For
 13 example, the USA was found to have a national food footprint of 1.0 ha/person, but a HALF of 1.8
 14 ha/person. The first value addresses, “how much land is used to produce the food consumed in the
 15 USA?”, and the second “how much land would be used if the global population adopted the average
 16 diet in the USA”. The inclusion of production systems within the land footprint to some degree
 17 obscures the understanding of the role of diet in the global food system. HALF, therefore, provides
 18 both a clearer comparative metric between countries of the land requirements of different diets, and
 19 also a way to consider the impacts from changes in dietary patterns.

20
 21 *(b) Temporal trends*
 22 Calculating the time-dependent HALF index for dietary variations only, i.e. assuming a constant 2011
 23 population and production systems, demonstrates the impacts of changes in food consumption
 24 patterns (solid lines in Figure 2). The global agricultural land required has increased by 8.7% due to
 25 dietary changes, from a HALF value of 32.3 in 1961 to 35.1 in 2011. For country-level average diets,
 26 results for Brazil and China show particularly substantial increases, due to the transitions in diets that
 27 are associated with increasing per capita wealth (Godfray et al., 2010), as well as the influence of
 28 urbanisation (Dong and Fuller, 2010; Huang and David, 1993; Popkin et al., 1999; Seto and Ramankutty,
 29 2016) and globalisation of food markets (Meyfroidt et al., 2013; Popkin, 2006). The land required for
 30 the diet in Brazil more than doubled between 1961 and 2011, from 43.5 to 88.2, making it the eleventh
 31 highest ranked country globally in 2011. However, the Chinese diet’s HALF increased nearly 5-times,
 32 from 6.0 in 1961 (the lowest at that period), to 28.6 (but still below the global average). The gap

1 between the USA and Indian diets has reduced slightly, from the USA value being 7.5 times the Indian
 2 value in 1961 to 6.2 times in 2011, with an 8% reduction in the USA and a 11% increase for the Indian
 3 diet.
 4



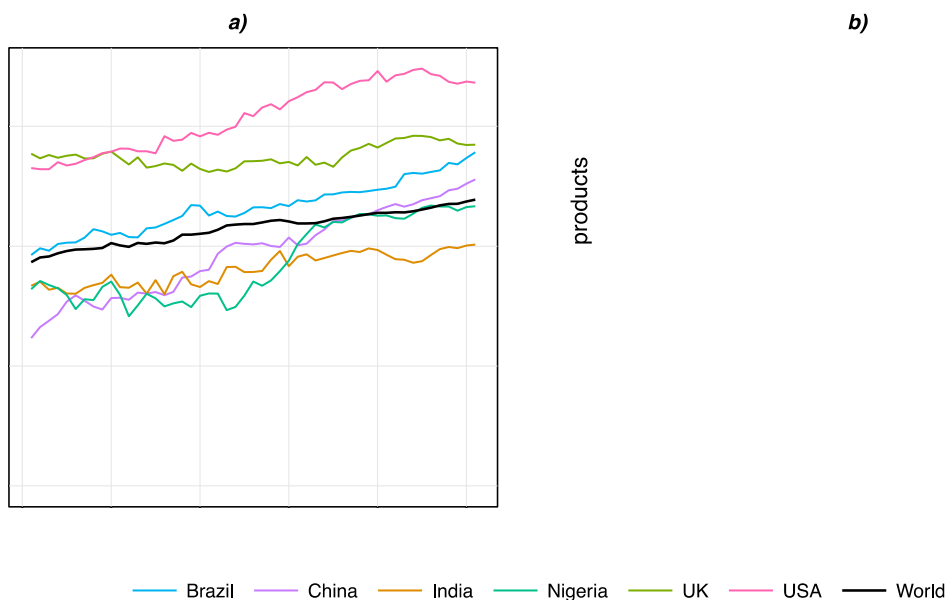
5
 6 *Figure 2. HALF index from 1961 to 2011, globally and for selected countries. Solid lines show variable*
 7 *diets, but constant population and agricultural production systems (at 2011 values). Dashed lines show*
 8 *variable diet, population and agricultural production systems over time.*

9
 10 When the time-dependent HALF indices are re-calculated to take account of changing production
 11 efficiencies and population sizes (Figure 2, dashed lines), they show a high degree of similarity to the
 12 diet-only case (Figure 2, solid lines). This is because increasing agricultural efficiencies and population
 13 growth in the past have acted in opposite directions on land requirements, largely offsetting one
 14 another. If production efficiencies from 2011 had been available and used in 1961, less than half of the
 15 agricultural land used at the time would have been required to feed the population at the time (Figure
 16 S2, dot-dashed line). However, populations have more than doubled since 1961, and therefore the
 17 2011 population would have required more than twice the land for food production based on 1961
 18 production systems (Figure S2, dotted line). The net effect is that if the mean global diet of 1961 had
 19 been consumed by the 2011 population, using 2011 production systems, agricultural land area would
 20 have remained largely unchanged from 1961 (just 5 Mha less land is estimated to have been needed
 21 than was used in 1961). When HALF values including variation in the production system and population
 22 (dashed lines in Figure 2) are lower than the HALF values for dietary changes only (solid lines), then
 23 cumulative improvements in agricultural efficiencies achieved by 2011 have not fully offset the rise in
 24 population. However, diets have also been changing. Dietary changes alone between 1961 and 2011
 25 has caused the agricultural area for food to increase by 368 Mha or 2.8% of the land surface. HALF has
 26 increased less than the 464 Mha expansion of global agricultural land since 1961 (FAOSTAT, 2015a), as

1 an increasing proportion of land is used for non-food uses of agricultural commodities, i.e. feedstocks
2 for biofuels.

3

4 The central role of the types of foods consumed in determining the agricultural land requirements of
5 different diets, compared to the overall quantity of nutrients consumed, can be seen from the
6 calculated energy intake and the percentage derived from animal products (Figure 3). Variation in total
7 food energy consumed between countries and over time is substantially smaller than the variations in
8 the land needed (Figure 3 & Figure S2). In 2011, the per capita land required to sustain a USA diet was
9 635% of that required for an Indian diet, even though the energy content of the food was only 65%
10 greater (or 99% greater in terms of protein; see Figure S3). This disparity stems from the profile of
11 commodities consumed, with 30% of energy derived from animal products in the USA and 9% in India
12 (65% and 19% respectively for protein). This greater proportion of animal products increases the land
13 requirements in comparison to a predominantly vegetarian diet, e.g. as in India.
14



15

16 *Figure 3. Mean energy per capita, a), and percentage energy derived from animal products, b), in foods*
17 *consumed from 1961 to 2011 globally, and for selected countries, using global average nutritional*
18 *values (FAOSTAT, 2015e, 2015f). This includes commodities wasted after reaching the consumer, but*
19 *not in the food supply chain.*

20

21 In developed countries such as the USA and the UK, per capita dietary land requirements have been
22 falling (Figure 2) even while energy and protein consumption continue to rise (Figure 3a & Figure S3a).
23 This apparent discrepancy is explained by the fall in the proportion of nutrients from animal products
24 (Figure 3b & Figure S3b), and a shift in the mix of animal products consumed (Figure 4). The drop in the
25 proportion of nutrients from animal products is in large part due to the increased consumption of
26 vegetal products, particularly vegetal oil, e.g. soybean oil. For example, in the USA vegetal oils provided
27 9.6% of calories in 1961, but this expanded to 19.2% by 2011 (14.5% from soya bean oil alone).
28 Consumption of these oils accounts for over half (55%) of the 3.2 MJ/person/day increase in energy
29 consumed in the USA, with other sweeteners (i.e. corn syrup) and poultry meat respectively accounting
30 for 26% and 18% of the rise.

31

32 The relative quantities of different animal products consumed changes over time, influencing the HALF
33 results. The effects of this are evident in the results for China, where since 1961 the proportion of

1 nutrients derived from animal products has increased towards that found in developed countries
 2 (Figure 3), but the HALF values have converged more slowly (Figure 2). The energy and protein intake
 3 and the percentages derived from animals are all higher than the global averages in China in 2011
 4 (Figure 3 & Figure S3). Nonetheless, the HALF is lower in China compared to its global value (Figure 2).
 5 This is due to the high rates of consumption of the commodities derived from monogastric animals
 6 (Figure 4), which have lower feed conversion ratios and lower land requirements in comparison to
 7 ruminants, although direct energy inputs are higher (Table 1). For example, the average diet in China
 8 contained around half the global average amount of beef (53%), but more than twice that of pork
 9 (239%). The rise in global HALF (8.5%) is also modest (Figure 2), given the rise in nutrients (28% rise in
 10 energy and protein) and the proportions derived from animals (increased by 11% for energy and 25%
 11 for protein). Again this can be understood by reference to the changes in the relative quantities of
 12 meats consumed (Figure 4). Global consumption per capita of bovine meat has been broadly constant,
 13 while poultry and pig meat have seen substantial rises, with 399% and 91% increases respectively from
 14 1961 to 2011. Global average per capita consumption of beef is now less than pork and poultry in
 15 mass, energy and protein.
 16



17
 18 *Figure 4. Per capita daily rates of bovine, pig and poultry meat consumption from 1961 to 2011. Data*
 19 *source: (FAOSTAT, 2015e).*

20
 21 *(c) Alternative diet scenarios*

22 Changes in diets and dietary impacts on land use are uncertain and are influenced by multiple factors,
 23 both economic and environmental. Two contrasting alternative scenario were used as exemplars to
 24 analyse the impacts of diet on global agricultural land use; the global adoption of the current diets of
 25 India and the USA. Although these countries are not the most extreme cases, they are major
 26 economies, with large populations, in which diets lie close to the lowest and highest land use
 27 requirements respectively (of the 170 countries included, India has the 13th lowest HALF value and the
 28 USA has the 6th highest, Table S5). Consideration of the adoption of these diets by the global
 29 population therefore provides a broad envelope within which human appropriation of land for food is
 30 likely to vary, but these are intended to be illustrative rather than represent equally plausible
 31 alternative futures. The net change in land use from a shift in global diet was decomposed into two
 32 parts; one considering a change in the quantity of nutrients consumed, and a second the profile of
 33 commodities consumed. The profile of commodities (i.e. the sources from which nutrients are derived)
 34 was found to have a greater impact on land use than the quantities of nutrients consumed, in the

1 dietary transitions considered (Table 2). For both dietary scenarios, changes in quantities and profiles
 2 act in the same direction, intensifying the overall impact.

3

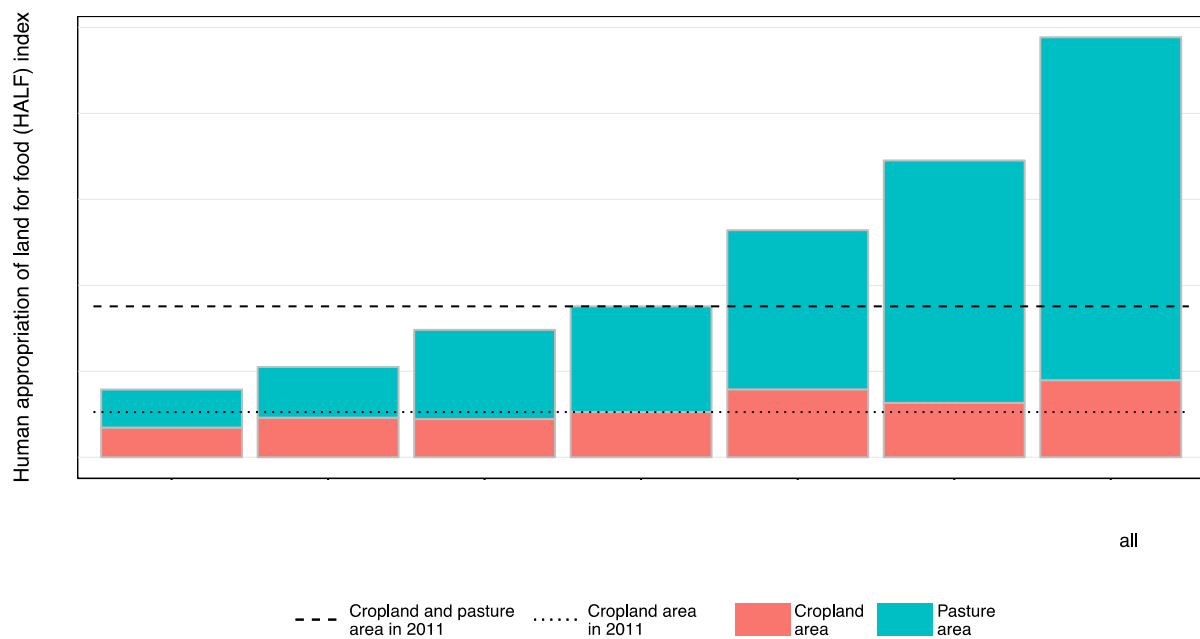
4 *Table 2. Changes in HALF from transitions of average global diet to that of India or the USA in 2011,*
 5 *divided into the impact from quantity of consumption ('quantity') and the types of commodities*
 6 *consumed ('profile'). For the quantity and profile cases, the change in areas are calculated based on*
 7 *providing the same energy and protein as current consumption. The overall type includes changes in*
 8 *quantities and profile of foods consumed, and by definition (1+overall change rate) = (1+profile change*
 9 *rate) * (1+quantity change rate), in terms of energy or protein. A single "overall" row is given for each*
 10 *dietary scenario, as this is equal in both nutrient terms.*

Dietary scenario country	Type and nutrient basis	Cropland area for food change (%)	Total cropland area change (%)	Livestock (feed & pasture) area change (%)	Agricultural area change (%)
India	Profile: Energy	+13	-22	-61	-47
India	Profile: Protein	+27	-12	-56	-40
India	Quantity: Energy	-16			
India	Quantity: Protein	-25			
India	Overall	-5	-34	-67	-55
USA	Profile: Energy	-11	+21	+122	+97
USA	Profile: Protein	-17	+13	+109	+85
USA	Quantity: Energy	+41			
USA	Quantity: Protein	+50			
USA	Overall	+25	+71	+214	+178

11

12 The impact of contrasting diets is much larger for the livestock area compared to cropland area used for
 13 food for human consumption. A more than 3-fold increase is required in livestock area (pasture and
 14 cropland for feed) under the USA diet scenario, increasing HALF by 178%. This area is needed both to
 15 support the increased quantities of nutrients consumed and the changes in dietary profile towards a
 16 greater proportion of animal products. Conversely, the lower overall consumption and the lower
 17 proportion from animal products in India suggests the livestock area would drop to less than a third of
 18 the current area, and reduce the overall HALF by 55%. The changes in cropland required to produce
 19 food for human consumption are comparatively modest with both the Indian and USA diets, with a 4%
 20 fall and a 21% rise respectively. The profile of the Indian diet is weighted towards vegetal crops, but
 21 the impact of this is offset by the lower level of nutrient intake overall. The opposite is the case for the
 22 average diet in America, with lower emphasis on crops, but higher overall consumption. Figure 5 shows
 23 the 2011 HALF index values for these scenarios, with cropland (for food and feed) and pasture
 24 identified separately.

25



1
 2 *Figure 5. Cropland and pasture required to produce food under alternative dietary scenarios, expressed*
 3 *as required percentage of world land, or HALF index, using global 2011 population and production*
 4 *systems. For each scenario (from Table 2) the case are shown that provides at least equal amounts of*
 5 *both energy and protein, e.g. the protein case is shown for the Indian diet profile, as the energy case*
 6 *provides insufficient protein.*

7

8 **4. Discussion**

9 *(a) Comparisons to previous studies*

10 The results show that global adoption of diets already consumed by hundreds of millions of people
 11 could lead to a magnitude of change greater than a doubling or halving of current agricultural land area.
 12 There have been few previous studies that have quantified the impact of such substantial shifts in diets
 13 on agricultural land areas. Stehfest *et al.* (2009) is one example, where dietary scenarios for 2050 are
 14 considered, including a ‘healthy diet’ (low rates of ruminant meat and pork and moderate poultry and
 15 consumption) and a no-meat diet. The current diet in India falls between these scenarios (i.e. rates of
 16 animal product consumption are lower than the Stehfest *et al.* ‘healthy diet’, but higher than the no-
 17 meat diet), and likewise the land use results found here lie between those of Stehfest *et al.* (2009). The
 18 impact of a ‘healthy diet’ was also considered in Bajželj *et al.* (2014), and showed a somewhat lower
 19 drop of 32% in pasture areas in 2050 compared to the authors’ business-as-usual scenario. The few
 20 studies published to date have shown that shifts in dietary preferences have a substantial impact not
 21 only on agricultural land use, but also on externalities such as GHG emissions and bioenergy potential
 22 (Haberl *et al.*, 2011; Popp *et al.*, 2010). Further studies that do not include land use change have also
 23 shown substantial GHG emissions implications from alternative diets, e.g. a 55% reduction from a
 24 vegetarian diet (Tilman and Clark, 2014). Considering the trade-offs between land for bioenergy
 25 production or afforestation (Williamson, 2016), reducing agricultural GHG emissions and meeting the
 26 food requirements of a growing population, a greater focus is justified in examining demand side
 27 measures, including waste reduction (Smith and Gregory, 2013).

28

29 The impact of global dietary changes since 1961 found here (Figure 2) is lower than that previously
 30 published (Alexander *et al.*, 2015). The differences arise primarily from the alternative approaches to
 31 allocating areas of monogastrics livestock. In Alexander *et al.* (2015) poultry and pigs were allocated a

1 proportion of pasture area, which increases the land use associated with these products, and
2 conversely reduces the ruminant products' footprint. However, monogastrics' nutrient requirements
3 are met from feed, while ruminants can also consume grass-based forage (Bellarby et al., 2013; Schader
4 et al., 2015). Therefore, in this study a more accurate assumption was made where only ruminants are
5 allocated a proportion of pasture area. As dietary changes have included larger increases in
6 monogastrics (than ruminant) derived productions (Figure 4) the resulting bias in Alexander *et al.*
7 (2015) associates dietary change with a greater land use impact than that found here. In 2011, 37.8%
8 of the world surface was used for agricultural purposes (FAOSTAT, 2015a), and here 34.5% was found
9 to be associated with food production. The difference between these rates is due to the other non-
10 food uses of agricultural commodities, such as bioenergy and fibre (Alexander et al., 2015; Rulli et al.,
11 2016).

12

13 (b) *Uncertainties in the analysis*

14 The results presented are derived under a set of assumptions with related uncertainties. Domestic
15 consumption is assumed to be supplied from the global production system. For example, countries
16 where grass-fed beef production systems predominate are treated identically to countries where
17 housed or feed-based systems are more common, as all use global average values. The distribution of
18 high HALF index values (Figure 1), appear to be associated with countries with substantial grassland
19 areas and high levels of beef production. This is not due directly to the production system, but to
20 these countries having high levels of beef consumption. The same effect occurs with vegetal
21 commodities, where countries with high production intensities and yields are assigned the same global
22 average as lower-yielding countries. Consequently, in countries with above-average yields, the HALF
23 areas associated with growing that crop would be higher than domestic production implies. The
24 national agricultural land footprints (Figure S1), gives the results of a similar calculation, but based on
25 domestic production and accounting for international trade (rather than a global average). Given the
26 research aims, we believe the approach of using a global average production systems is reasonable
27 because of the global scale of the analysis (considering global adoption of alternative diets), and also
28 because of the levels of international trade in agricultural commodities and the associated globalised
29 markets (D'Odorico et al., 2014; Fader et al., 2013; Meyfroidt et al., 2013). Most importantly, the
30 approach allows the impact of variations in diets to be quantified without the obscuring influence of
31 differences in the production system.

32

33 The disaggregation of feed by animal products uses the feed requirements calculated from feed
34 conversion ratios (FCR; Table 1). FCR are difficult to estimate, and have been the subject of
35 misrepresentation by both sides of the sustainability - meat consumption debate (Fairlie, 2010). The
36 FCRs used here are for the global average production, derived in FAO studies (Macleod et al., 2013;
37 Opio et al., 2013). While some uncertainty in FCRs remains, changes in the ratios only affect the
38 disaggregation of the global pasture and feed areas between animal products. Biases introduced by
39 inaccurate FCRs will cancel out in the baseline case. When alternative consumption profiles are
40 considered they may not perfectly cancel out, and result in a residual bias in the required land areas
41 calculated. This is likely to be small relative to the scale of the overall effects shown, due in part to the
42 offsetting between animal products. As a check on the accuracy of the FCRs used, the allocation of feed
43 between monogastric animal and ruminants was compared against the results of a survey of the feed
44 use from 134 countries (Alltech, 2013). This survey showed that 26% of total feed use was for
45 ruminants in 2012, while 23% of feed was calculated as used for ruminants in 2011 in the results
46 presented here. The level of agreement between these values gives additional confidence in the FCR
47 rates used.

48

1 (c) *Obesity, malnutrition and waste*

2 The findings presented here are based on the average food reaching consumers rather than human
3 nutritional requirements, and it is important to consider the extent to which these differ within a
4 population. Distinctions arise due to over-eating and, conversely, malnutrition, through waste of food
5 by consumers (Eshel and Martin, 2006), and also inequalities in distribution (Porkka et al., 2013).
6 Losses and waste occur at each stage of the food supply chain, with overall food waste, accounting for
7 losses in production and at the consumer, estimated to be around 25-40% of total food production
8 (Godfray et al., 2010; Kummu et al., 2012). HALF values include losses both in the production system
9 (e.g. unharvested crops and losses in storage, transportation, and processing) and at the consumer.
10 Production system losses are derived from the global production efficiencies, and therefore are
11 considered only as a global average. By contrast, food waste by consumers are included at a country
12 specific level, as this is included in the FAO commodity balance data used (FAOSTAT, 2015d).
13 Consequentially, the HALF index includes (but does not separately identify) the variations in the rates of
14 per capita food waste by consumers. 95-115 kg/year of food has been estimated to be wasted per
15 capita after reaching the consumer in Europe and North-America, while in sub-Saharan Africa and
16 South/Southeast Asia this is only 6-11 kg/year (Gustavsson et al., 2011), which equates to 9-12% and 1-
17 3% of food delivered to consumers respectively. Applying the mean values of these rates for USA and
18 India suggests that the HALF values for consumer wastes alone is 10.3 and 0.3, respectively.

19
20 The protein requirement of adult men and women depends on body weight. For an average body
21 weight of 60kg, 50 g/day of protein is the minimum safe limit (WHO et al., 2007). No country with a
22 population of more than 20 million currently falls below this limit, although several smaller countries
23 consume 40-50 g/person/day, i.e. Guinea, Guinea-Bissau, Haiti, Liberia, Madagascar, Mozambique,
24 Zambia and Zimbabwe. The energy requirements also vary by sex, weight and the level of physical
25 activity. For instance, average energy requirements for the population of UK adult females and males,
26 are respectively 8.7 MJ/day (2079 kcal/day) and 10.9 MJ/day (2605 kcal/day) (SACN, 2011). To
27 compare with the calculated energy in-takes, we assume the mean energy requirement value is 9.8
28 MJ/person/day (2342 kcal/person/day). This value is somewhat higher than the 2100 kcal/person/day
29 energy intake used in some previous studies (Eshel and Martin, 2006; Kummu et al., 2012), and likely to
30 exceed the in-take needed to avoid hunger or malnutrition (WFP, 2016). The average Indian
31 consumption appears close to the population's energy requirements, given the relatively low levels of
32 consumer waste in South & Southeast Asia (Gustavsson et al., 2011), just 1% more, assuming 2% food is
33 discarded.

34
35 Even if there is sufficient food to avoid malnutrition within a country or region, this does not mean that
36 these foods are distributed equitably. Globally, 37% of men and 38% of women were overweight in
37 2014 (Ng et al., 2014), while approximately 12% of people were undernourished between 2010 and
38 2012 (FAO et al., 2015). The populations living in countries with critically low food supply (<2000
39 kcal/cap/d) has also been dropping over time, from 52% in 1965 to 3% in 2005 (Porkka et al., 2013). In
40 India (ranked 25th worst in the 2015 Global Hunger Index Report (Grebmer et al., 2015)) 20% of the
41 population are over-weight (including nearly 5% obese) and 15% undernourished (FAO et al., 2015; Ng
42 et al., 2014), while the for adults in the USA 66% are over-weight, including 33% obese (Ng et al., 2014).
43 Given there are three-times more overweight people than undernourished, and that levels of
44 malnutrition have been declining over recent years, better national and international distribution of
45 food is more relevant to achieving global food security than additional production.

46
47 The USA per capita energy consumption is 16.6 MJ/day, which suggests that 41% of food (in energy
48 terms) is either due to overeating or consumer waste (34% of energy intake is in excess of
49 requirements, assuming 10.5% food waste (Gustavsson et al., 2011)). This is in line with a previous

1 finding, showing that in the USA, overeating and food discarded by consumers accounted for 44% of
2 food distributed to consumers (Eshel and Martin, 2006). The results suggest that under the global
3 adoption of USA consumer behaviours the land required to produce the food wasted by consumers
4 (including over-consumption), would be sufficient to provide more than twice the entire food
5 requirements assuming adoption of Indian consumption patterns.

6 7 *(d) Plausibility of dietary scenarios*

8 Two contrasting scenarios were used to examine how changes in food consumption preferences and
9 behaviours might affect agricultural commodity demand and land use. These scenarios explore the
10 consequences of a wide range of consumption patterns, but do not represent equally plausible future
11 states. The first scenario considers the average global diet transitioning to the current average USA
12 diet. Although this (time-independent) scenario is unlikely in the short term, consumption patterns
13 have been shifting in this direction, due to increases in per capita incomes in developing countries (e.g.
14 China and Brazil), rural-urban migration and globalisation, leading to more overall per capita food
15 consumption, and a greater percentage consumption of animal products (Lambin and Meyfroidt, 2011;
16 Seto and Ramankutty, 2016; Tilman et al., 2011). However, a substantial gap in consumption patterns
17 remains between countries, with the US diet requiring 2.8 times the land area of the global average
18 diet, and 3.4 times that of the Chinese diet. Consequently, given current yields and production
19 systems, it would clearly not be possible for the world's population to consume food as in the US;
20 indeed, this would require 98% of all land, including snow-cover and deserts. Apart from being
21 physically impossible, changes to approach this level of consumption would also generate strong
22 market signals that would act to increase the price of food, suppress demand and intensify production
23 practices (additional inputs, e.g. irrigation water, fertiliser or labour, leading to higher yield).
24 Conversely, if more land were to be used for agriculture, suitable land would become more scarce, and
25 the additional land would tend to be of lower quality and produce lower yields, leading to a greater
26 area requirements (Lambin and Meyfroidt, 2011). Price signals may be particularly large for the less
27 efficient and potentially costlier commodities, e.g. beef. Arguably, these impacts are already evident,
28 with a shift towards chicken and away from beef (Figure 4) supported by intensification of chicken
29 production and the associated efficiency increases (Havenstein, 2006).

30
31 The contrasting scenario considers the global diet becoming equivalent to the average diet of India.
32 This is more plausible from an environmental and agricultural system viewpoint. However, it implies
33 shifts in consumption that are the opposite of the global consumption trends that have occurred over
34 previous decades, as per capita incomes have increased in developing countries. A reversal of these
35 trends would either require a substantial shift in consumer preferences (towards the consumption of
36 vegetal crops, e.g. higher rates of vegetarianism), or a catastrophic global economic collapse reducing
37 per capita incomes, particularly in wealthier countries. Changes in food preferences may be achievable
38 through either behavioural or economic approaches. For example, less food is consumed when people
39 are offered smaller-sized portions, packages or tableware than when offered larger-sized versions,
40 leading to the possibility of policies to reduce consumption (Hollands et al., 2015). Economic
41 approaches such as taxes (e.g. a fat tax or a tax on sugar-sweetened beverages) and subsidies (e.g. on
42 fruit and vegetables) could be used to provide fiscal incentives to change behaviours (Thow et al., 2010;
43 Wang et al., 2012). However, the effectiveness of taxation and subsidies alone to alter diets, without
44 other policies that target a number of different levels within society, has been questioned (Tiffin and
45 Arnoult, 2011).

5. Conclusions

Dramatically different requirements for land for food production could arise depending on the course of dietary change – both in terms of quantity of food consumed per person, but more importantly in terms of the mix of food commodities. A wide range of human appropriation of land for food was found based on global adoption of current country-level average diets, far wider than the divergence in energy or protein in-takes, with the difference due to the types of commodities in each diet, and in particular the level of ruminant animal products. For example, if the diets of India or the USA were adopted globally the impact from the change in the mix of commodities would be about twice that from the quantities consumed. What we individually eat (or even waste), rather than how much, appears to be more important for agricultural land requirements. However, waste and over-eating are still important issues, with the results suggesting that the land required to produce the food wasted by consumers (including over-consumption) given USA consumption, could provide more than twice the food required under adoption of Indian consumption patterns.

Shifts toward diets of Western countries, exemplified here by the average diet in the USA, for the global population are not sustainable or desirable for environmental and health reasons (Tilman and Clark, 2014). Given the possibility that intensification alone may be insufficient to satisfy changes in dietary preferences and population growth, other methods of avoiding increases in agricultural areas are needed to target consumer behaviours or preferences. Behavioural and economic mechanisms need to be better understood to establish how more equitable, healthy and environmentally benign food consumption can be achieved.

6. References

- Alexander, P., Rounsevell, M.D.A., Dislich, C., Dodson, J.R., et al., 2015. Drivers for global agricultural land use change: The nexus of diet, population, yield and bioenergy. *Global Environmental Change* 35, 138–147. doi:10.1016/j.gloenvcha.2015.08.011
- Alltech, 2013. *Global Feed Summary*. Alltech, Nicholasville, Kentucky, USA.
- Bajželj, B., Richards, K.S., Allwood, J.M., Smith, P., et al., 2014. Importance of food-demand management for climate mitigation. *Nature Climate Change* 4, 924–929. doi:10.1038/nclimate2353
- Bellarby, J., Tirado, R., Leip, A., Weiss, F., et al., 2013. Livestock greenhouse gas emissions and mitigation potential in Europe. *Global Change Biology* 19, 3–18. doi:10.1111/j.1365-2486.2012.02786.x
- Carlsson-Kanyama, A., González, A.D., 2009. Potential contributions of food consumption patterns to climate change. *The American journal of clinical nutrition* 89, 1704S–1709S. doi:10.3945/ajcn.2009.26736AA.1704S
- Cassidy, E.S., West, P.C., Gerber, J.S., Foley, J. a, 2013. Redefining agricultural yields: from tonnes to people nourished per hectare. *Environmental Research Letters* 8, 34015. doi:10.1088/1748-9326/8/3/034015
- D’Odorico, P., Carr, J. a., Laio, F., Ridolfi, L., Vandoni, S., 2014. Feeding humanity through global food trade. *Earth’s Future* 2, 458–469. doi:10.1002/2014EF000250
- Davis, K.F., Odorico, P.D., Rulli, M.C., 2014. Moderating diets to feed the future. *Earth’s Future* 2, 559–565. doi:10.1002/2014EF000254.Received
- Davis, K.F., Yu, K., Herrero, M., Havlik, P., et al., 2015. Historical trade-offs of livestock’s environmental impacts. *Environmental Research Letters* 10, 125013. doi:10.1088/1748-9326/10/12/125013
- Dong, F., Fuller, F., 2010. Dietary structural change in China’s cities: Empirical fact or urban legend? *Canadian Journal of Agricultural Economics* 58, 73–91. doi:10.1111/j.1744-7976.2009.01159.x
- Engström, K., Rounsevell, M.D.A., Murray-Rust, D., Hardacre, C., et al., 2016. Applying Occam’s Razor to global agricultural land use change. *Environmental Modelling & Software* 75, 212–229. doi:10.1016/j.envsoft.2015.10.015

1 Erb, K.-H., Krausmann, F., Lucht, W., Haberl, H., 2009. Embodied HANPP: Mapping the spatial
2 disconnect between global biomass production and consumption. *Ecological Economics* 69, 328–
3 334. doi:10.1016/j.ecolecon.2009.06.025

4 Eshel, G., Martin, P.A., 2006. Diet , Energy , and Global Warming. *Earth Interactions* 10, 1–17.

5 Fader, M., Gerten, D., Krause, M., Lucht, W., Cramer, W., 2013. Spatial decoupling of agricultural
6 production and consumption: quantifying dependences of countries on food imports due to
7 domestic land and water constraints. *Environmental Research Letters* 8, 14046. doi:10.1088/1748-
8 9326/8/1/014046

9 Fairlie, S., 2010. Meat: A benign extravagance. Permanent Publications, East Meon, Hampshire, UK.

10 FAO, 2006. Livestock's long shadow - environmental issues and options. Food and Agriculture
11 Organization of the United Nations (FAO), Rome, Italy. doi:10.1007/s10666-008-9149-3

12 FAO, IFAD, WFP, 2015. The State of Food Insecurity in the World: Meeting the 2015 international
13 hunger targets: taking stock of uneven progress. Food and Agriculture Organization of the United
14 Nations (FAO), Rome, Italy. doi:14646E/1/05.15

15 FAOSTAT, 2015a. Resources/Land (2015-12-16). Food and Agriculture Organization of the United
16 Nations, Rome, Italy.

17 FAOSTAT, 2015b. Commodity Balances/Livestock and Fish Primary Equivalent (2015-12-16). Food and
18 Agriculture Organization of the United Nations, Rome, Italy.

19 FAOSTAT, 2015c. Production/Crops (2015-12-16). Food and Agriculture Organization of the United
20 Nations, Rome, Italy.

21 FAOSTAT, 2015d. Commodity Balances/Crops Primary Equivalent (2015-12-16). Food and Agriculture
22 Organization of the United Nations, Rome, Italy.

23 FAOSTAT, 2015e. Food Supply - Livestock and Fish Primary Equivalent (2015-12-16). Food and
24 Agriculture Organization of the United Nations, Rome, Italy.

25 FAOSTAT, 2015f. Food Supply - Crops Primary Equivalent (2015-12-16). Food and Agriculture
26 Organization of the United Nations, Rome, Italy.

27 FAOSTAT, 2015g. Population/Annual time series (2015-12-16). Food and Agriculture Organization of the
28 United Nations, Rome, Italy.

29 FAOSTAT, 2015h. Production/Livestock Primary (2015-12-16). Food and Agriculture Organization of the
30 United Nations, Rome, Italy.

31 Foley, J.A., Ramankutty, N., Brauman, K.A., Cassidy, E.S., et al., 2011. Solutions for a cultivated planet.
32 *Nature* 478, 337–42. doi:10.1038/nature10452

33 Godfray, H.C.J., Beddington, J.R., Crute, I.R., Haddad, L., et al., 2010. Food security: the challenge of
34 feeding 9 billion people. *Science (New York, NY)* 327, 812–8. doi:10.1126/science.1185383

35 González, A.D., Frostell, B., Carlsson-Kanyama, A., 2011. Protein efficiency per unit energy and per unit
36 greenhouse gas emissions: Potential contribution of diet choices to climate change mitigation.
37 *Food Policy* 36, 562–570. doi:10.1016/j.foodpol.2011.07.003

38 Grebmer, K. von, Bernstein, J., Prasai, N., Yin, S., Yohannes, Y., 2015. 2015 Global Hunger Index.
39 International Food Policy Research Institute, Bonn, Washington, DC, and Dublin.

40 Gustavsson, J., Cederberg, C., Sonesson, U., Otterdijk, R. van, Meybeck, A., 2011. Global food losses and
41 food waste– Extent, causes and prevention. Food and Agriculture Organization of the United
42 Nations (FAO), Rome, Italy.

43 Haberl, H., Erb, K.H., Krausmann, F., Bondeau, A., et al., 2011. Global bioenergy potentials from
44 agricultural land in 2050: Sensitivity to climate change, diets and yields. *Biomass and Bioenergy*
45 35, 4753–4769. doi:10.1016/j.biombioe.2011.04.035

46 Havenstein, G.B., 2006. Performance changes in poultry and livestock following 50 years of genetic
47 selection. *Lohmann Information* 41, 30–37.

48 Herrero, M., Conant, R., Havlik, P., Hristov, A.N., et al., 2016. Greenhouse gas mitigation potentials in
49 the livestock sector. *Nature Climate Change*. doi:10.1038/nclimate2925

50 Hollands, G., Shemilt, I., Marteau, T., Jebb, S., et al., 2015. Portion , package or tableware size for
51 changing selection and consumption of food , alcohol and tobacco. *Cochrane Database of*
52 *Systematic Reviews*. doi:10.1002/14651858.CD011045.pub2.Copyright

1 Hu, F.B., 2011. Globalization of Diabetes: The role of diet, lifestyle, and genes. *Diabetes Care* 34, 1249–
2 1257. doi:10.2337/dc11-0442

3 Huang, J., Bouis, H., 2001. Structural changes in the demand for food in Asia: Empirical evidence from
4 Taiwan. *Agricultural Economics* 26, 57–69. doi:10.1016/S0169-5150(00)00100-6

5 Huang, J., David, C.C., 1993. Demand for cereal grains in Asia: The effect of urbanization. *Agricultural*
6 *Economics* 8, 107–124. doi:10.1016/0169-5150(92)90025-T

7 Index Mundi, 2016. Commodity Prices Indices: Vegetable Oil and Protein Meal [WWW Document]. URL
8 <http://www.indexmundi.com/commodities/?commodity=soybean-oil> (accessed 1.7.16).

9 INRA, CIRAD, AFZ, FAO, 2016. Animal feed resources information system, Feedipedia.

10 Jalava, M., Kummu, M., Porkka, M., Siebert, S., Varis, O., 2014. Diet change—a solution to reduce water
11 use? *Environmental Research Letters* 9, 74016. doi:074016 10.1088/1748-9326/9/7/074016

12 Kastner, T., Erb, K.-H., Haberl, H., 2014. Rapid growth in agricultural trade: effects on global area
13 efficiency and the role of management. *Environmental Research Letters* 9, 34015.
14 doi:10.1088/1748-9326/9/3/034015

15 Kastner, T., Rivas, M.J.I., Koch, W., Nonhebel, S., 2012. Global changes in diets and the consequences
16 for land requirements for food. *Proceedings of the National Academy of Sciences of the United*
17 *States of America* 109, 6868–6872. doi:10.1073/pnas.1117054109

18 Krausmann, F., Erb, K.-H., Gingrich, S., Haberl, H., et al., 2013. Global human appropriation of net
19 primary production doubled in the 20th century. *Proceedings of the National Academy of Sciences*
20 110, 10324–10329. doi:10.1073/pnas.1211349110

21 Kummu, M., de Moel, H., Porkka, M., Siebert, S., et al., 2012. Lost food, wasted resources: Global food
22 supply chain losses and their impacts on freshwater, cropland, and fertiliser use. *Science of the*
23 *Total Environment* 438, 477–489. doi:10.1016/j.scitotenv.2012.08.092

24 Lambin, E.F., Meyfroidt, P., 2011. Global land use change, economic globalization, and the looming
25 land scarcity. *Proceedings of the National Academy of Sciences of the United States of America*
26 108, 3465–3472. doi:10.1073/pnas.1100480108

27 Le Quéré, C., Moriarty, R., Andrew, R.M., Peters, G.P., et al., 2015. Global carbon budget 2014 47–85.
28 doi:10.5194/essd-7-47-2015

29 Little, S., 2014. Feed Conversion Efficiency: A key measure of feeding system performance on your
30 farm. Dairy Australia, Victoria, Australia.

31 Macleod, M., Gerber, P., Mottet, A., Tempio, G., et al., 2013. Greenhouse gas emissions from pig and
32 chicken supply chains - A global life cycle assessment. Food and Agriculture Organization of the
33 United Nations (FAO), Rome, Italy.

34 Marlow, H.J., Hayes, W.K., Soret, S., Carter, R.L., et al., 2009. Diet and the environment: does what you
35 eat matter? *American Journal of Clinical Nutrition* 89, 1699S–1703S.
36 doi:10.3945/ajcn.2009.26736Z

37 Meyfroidt, P., Lambin, E.F., Erb, K.-H., Hertel, T.W., 2013. Globalization of land use: distant drivers of
38 land change and geographic displacement of land use. *Current Opinion in Environmental*
39 *Sustainability* 5, 438–444. doi:10.1016/j.cosust.2013.04.003

40 Mueller, N.D., Gerber, J.S., Johnston, M., Ray, D.K., et al., 2012. Closing yield gaps through nutrient and
41 water management. *Nature* 490, 254–7. doi:10.1038/nature11420

42 Ng, M., Fleming, T., Robinson, M., Thomson, B., et al., 2014. Global, regional, and national prevalence of
43 overweight and obesity in children and adults during 1980–2013: a systematic analysis for the
44 Global Burden of Disease Study 2013. *The Lancet* 384, 766–781. doi:10.1016/S0140-
45 6736(14)60460-8

46 Opio, C., Gerber, P., Mottet, A., Falculli, A., et al., 2013. Greenhouse gas emissions from ruminant
47 supply chains- A global life cycle assessment. Food and Agriculture Organization of the United
48 Nations (FAO), Rome, Italy.

49 Parfitt, J., Barthel, M., Macnaughton, S., 2010. Food waste within food supply chains: quantification and
50 potential for change to 2050. *Philosophical Transactions of the Royal Society B: Biological Sciences*
51 365, 3065–3081. doi:10.1098/rstb.2010.0126

52 Pelletier, N., Audsley, E., Brodt, S., Garnett, T., et al., 2011. Energy Intensity of Agriculture and Food

1 Systems. *Annual Review of Environment and Resources* 36, 223–246. doi:10.1146/annurev-
2 environ-081710-161014

3 Pingali, P., 2007. Westernization of Asian diets and the transformation of food systems: Implications for
4 research and policy. *Food Policy* 32, 281–298. doi:10.1016/j.foodpol.2006.08.001

5 Popkin, B.M., 2006. Technology, transport, globalization and the nutrition transition food policy. *Food*
6 *Policy* 31, 554–569. doi:10.1016/j.foodpol.2006.02.008

7 Popkin, B.M., Carolina, N., Hill, C., 1999. Popkin(1999) Urbanization, Lifestyle Changes and the Nutrition
8 27, 1905–1916.

9 Popp, A., Lotze-Campen, H., Bodirsky, B., 2010. Food consumption, diet shifts and associated non-CO2
10 greenhouse gases from agricultural production. *Global Environmental Change* 20, 451–462.
11 doi:10.1016/j.gloenvcha.2010.02.001

12 Porkka, M., Kummu, M., Siebert, S., Varis, O., 2013. From food insufficiency towards trade dependency:
13 A historical analysis of global food availability. *PLoS ONE* 8. doi:10.1371/journal.pone.0082714

14 Rulli, M.C., Bellomi, D., Cazzoli, A., Carolis, G. De, Odorico, P.D., 2016. The water-land-food nexus of
15 first- generation biofuels. *Nature Publishing Group* 1–10. doi:10.1038/srep22521

16 SACN, 2011. *Dietary Reference Values for Energy 2011*. Scientific Advisory Committee on Nutrition,
17 London, UK.

18 Schader, C., Muller, A., Scialabba, N.E.-H., Hecht, J., et al., 2015. Impacts of feeding less food-competing
19 feedstuffs to livestock on global food system sustainability. *Journal of The Royal Society Interface*
20 12, 20150891. doi:10.1098/rsif.2015.0891

21 Schmitz, C., van Meijl, H., Kyle, P., Nelson, G.C., et al., 2014. Land-use change trajectories up to 2050:
22 insights from a global agro-economic model comparison. *Agricultural Economics* 45, 69–84.
23 doi:10.1111/agec.12090

24 Seto, K.C., Ramankutty, N., 2016. Hidden linkages between urbanization and food systems. *Science* 352,
25 943–945.

26 Smil, V., 2013. *Should We Eat Meat? Evolution and Consequences of Modern Carnivory*. Wiley, New
27 York, USA.

28 Smil, V., 2002. Worldwide transformation of diets, burdens of meat production and opportunities for
29 novel food proteins. *Enzyme and Microbial Technology* 30, 305–311. doi:10.1016/S0141-
30 0229(01)00504-X

31 Smith, P., Gregory, P.J., 2013. Climate change and sustainable food production. *The Proceedings of the*
32 *Nutrition Society* 72, 21–8. doi:10.1017/S0029665112002832

33 Smith, P., Haberl, H., Popp, A., Erb, K.-H., et al., 2013. How much land-based greenhouse gas mitigation
34 can be achieved without compromising food security and environmental goals? *Global change*
35 *biology* 19, 2285–302. doi:10.1111/gcb.12160

36 Stehfest, E., Bouwman, L., Van Vuuren, D.P., Den Elzen, M.G.J., et al., 2009. Climate benefits of changing
37 diet. *Climatic Change* 95, 83–102. doi:10.1007/s10584-008-9534-6

38 Thow, A.M., Jan, S., Leeder, S., Swinburn, B., 2010. The effect of fiscal policy on diet, obesity and
39 chronic disease: a systematic review. *Bulletin of the World Health Organization* 88, 609–614.
40 doi:10.2471/BLT.09.070987

41 Tiffin, R., Arnoult, M., 2011. The public health impacts of a fat tax. *European Journal of Clinical Nutrition*
42 65, 427–433. doi:10.1038/ejcn.2010.281

43 Tilman, D., Balzer, C., Hill, J., Befort, B.L., 2011. Global food demand and the sustainable intensification
44 of agriculture. *Proceedings of the National Academy of Sciences of the United States of America*
45 108, 20260–4. doi:10.1073/pnas.1116437108

46 Tilman, D., Clark, M., 2014. Global diets link environmental sustainability and human health. *Nature*
47 515, 518–522. doi:10.1038/nature13959

48 van Vuuren, D.P., Carter, T.R., 2014. Climate and socio-economic scenarios for climate change research
49 and assessment: Reconciling the new with the old. *Climatic Change* 122, 415–429.
50 doi:10.1007/s10584-013-0974-2

51 Wang, Y.C., Coxson, P., Shen, Y., Goldman, L., 2012. A Penny-Per-Ounce Tax On Sugar-Sweetened
52 Beverages Would Cut Health And Cost Burdens Of Diabetes. *Health Affairs* 31, 199–207.

1 doi:10.1377/hlthaff.2011.0410
2 Weinzettel, J., Hertwich, E.G., Peters, G.P., Steen-Olsen, K., Galli, A., 2013. Affluence drives the global
3 displacement of land use. *Global Environmental Change* 23, 433–438.
4 doi:10.1016/j.gloenvcha.2012.12.010
5 West, P.C., Gerber, J.S., Engstrom, P.M., Mueller, N.D., et al., 2014. Leverage points for improving global
6 food security and the environment. *Science* 345, 325–328. doi:10.1126/science.1246067
7 WFP, 2016. What is hunger? World Food Programme (WFP), Rome, Italy.
8 WHO, FAO, UNU, 2007. Protein and amino acid requirements in human nutrition. World Health
9 Organization technical report series 935.
10 Williamson, P., 2016. Scrutinize CO 2 removal methods. *Nature* 530, 5–7.
11 Wu, Y., Wu, H.X., 1997. Household Grain Consumption in China : Effects of Income , Price and
12 Urbanization * Yanrui Wu. *Asian Economic Journal* 11, 325–342.
13 Yu, Y., Feng, K., Hubacek, K., 2013. Tele-connecting local consumption to global land use. *Global*
14 *Environmental Change* 23, 1178–1186. doi:10.1016/j.gloenvcha.2013.04.006
15
16

Supplementary Information: Human appropriation of land for food: the role of diet

SI Methods

(a) Example cropland area allocation calculations

Taking Soyabeans and its products as an example, the globally aggregated values for 2011 are shown in Table S1.

Table S1. Soyabean commodity balance data in 2011. Data source: (FAOSTAT, 2015d).

Commodity	Production (Mt)	Food (Mt)	Feed (Mt)	Processing (Mt)	Non-food (Mt)	Seed (Mt)	Waste (Mt)	Stock variation (Mt)
Soyabeans	261.9	10.1	13.3	224.1	0.7	6.8	4.0	-2.1
Soyabean Oil	41.8	24.6	0.0	0.0	16.7	0	0.2	-0.3
Soyabean Cake	176.7	0	171.6	0	3.2	0	0	-0.0

The area specified as used for growing soyabeans is 103.8 Mha (FAOSTAT, 2015c). The area calculated as being used to produce each category of use (food, feed, processing and non-food) is as per the following example:

Soyabean processing area

$$\begin{aligned}
 &= \text{soyabean production area} * \text{soyabean processing quantity} / (\text{soyabean food, feed, processing, non-food and stock variation}) * \text{total cropland area} / \text{total harvested area} \\
 &= 103.8 \text{ Mha} * 224.1 \text{ Mt} / (10.1 \text{ Mt} + 13.3 \text{ Mt} + 224.1 \text{ Mt} + 0.7 \text{ Mt} - (-2.1)) * 1556 \text{ Mha} / 1378 \text{ Mha} \\
 &= 104.9 \text{ Mha}
 \end{aligned}$$

Stock variation is subtracted as it is defined in the FAO dataset such that “net increases in stocks (add to stock) are generally indicated by the negative sign”.

The primary crop processing area is then mapped onto the commodities that result from the processing, with the allocation by the economic value of these resultant products. For oil crops is assumed to be 50:50 (see Table S4). Therefore, the production area for soyabean oil and soyabean cake are both 52.9 Mha. To obtain the areas associated with each use (i.e. food, feed, processing, non-food) these use the same approach as above. For example, area for Soyabean oil this is as follows:

Soyabean oil feed area

$$\begin{aligned}
 &= 52.5 \text{ Mha} * 24.6 \text{ Mt} / (24.6 \text{ Mt} + 0.0 \text{ Mt} + 0.0 \text{ Mt} + 16.7 \text{ Mt} - (-0.3)) \\
 &= 31.0 \text{ Mha}
 \end{aligned}$$

(b) Example animal feed and pasture area calculation

The feed requirements, assuming all nutrients are from feed, are calculated using the FCR and quantity of the livestock product produced. For monogastrics species these values are taken as the feed amounts, and the associated areas used to grow these feeds. For example in 2011 global, 102.5 Mt of poultry meat was produced (FAOSTAT, 2015h). A FCR of 3.3 (Table 1) implies that 338.2 Mt of feed DM is required for poultry meat production. The total feed DM specified in the commodity balance data (converted to DM using feed moisture contents (INRA et al., 2016)) is 1515.4 Mt produced from 504.4 Mha of cropland. Therefore, we can calculate the area for poultry meat feed as:

Poultry meat feed area

$$\begin{aligned} &= \text{Poultry meat feed requirements} * \text{total feed area} / \text{total feed quantity} \\ &= \text{Poultry meat produced} * \text{poultry FCR} * \text{total feed area} / \text{total feed quantity} \\ &= 102.5 \text{ Mt} * 3.3 * 504.4 \text{ Mha} / 1515.4 \text{ Mt} \\ &= 112.6 \text{ Mha} \end{aligned}$$

The same is applied for the other two monogastrics products, i.e. eggs and pig meat, giving feed requirements of 161.9 Mt and 690.9 Mt respectively. This implies that of the total 1515.4 Mt of feed 324.4 Mt is consumed by the ruminant species (plus all the pasture). The area used to produce the ruminant feed is 108.0 Mha ($324.4 \text{ Mt} * 504.4 \text{ Mha} / 1515.4 \text{ Mt}$). The ruminant feed and pasture areas are divided in the same manner (pro rata by feed requirement between ruminants), for example:

Bovine meat feed area

$$\begin{aligned} &= \text{Ruminant feed area} * \text{bovine meat feed requirements} / \text{total ruminant feed requirements} \\ &= \text{Ruminant feed area} * \text{bovine meat quantity produced} * \text{bovine meat FCR} / \text{sum for all} \\ &\text{ruminant (quantity produced} * \text{FCR)} \\ &= 108.0 \text{ Mha} * 66.2 \text{ Mt} * 25.0 / 2809 \text{ Mt} \\ &= 63.6 \text{ Mha} \end{aligned}$$

Bovine meat pasture area

$$\begin{aligned} &= \text{Pasture area} * \text{bovine meat feed requirements} / \text{total ruminant feed requirements} \\ &= 3364 \text{ Mha} * 66.2 \text{ Mt} * 25.0 / 2809 \text{ Mt} \\ &= 1982 \text{ Mha} \end{aligned}$$

Table S2. Summary of key assumptions used to calculate the HALF index.

Context	Assumption	Comment/implications
Supply	Production efficiency, including production losses, equals the global average in a baseline year of 2011. This assumption is varied in some analyses.	Fixed production system, so no change in intensity in response to shifts in demand. However, an increase in the HALF can imply an increase in agricultural areas, an increase in productive efficiency (e.g. intensity), or a combination of both. The global nature of the analysis is predicated upon the globalised agricultural commodity markets.
Supply	Global average production efficiencies apply for all countries.	
Supply	Agricultural production does not adapt to demand changes.	
Supply: Livestock	Mix of commodities used for animal feeds remains constant, and is set by the rates of use in the baseline data.	
Supply: Livestock	Monogastric livestock, i.e. pigs and poultry, nutritional requirements are derived exclusively from feed.	Disaggregation of cropland areas used to grow animal feed and pasture uses feed conversion ratios. Any inaccuracies in these values impacts the allocation of the areas between livestock product in the global case, but not the total areas.
Supply: Livestock	Ruminant livestock, e.g. cattle, consume remaining feed (after meeting monogastrics requirements) and all pasture in proportion to the ruminant livestock's nutritional requirements.	
Supply: Livestock	Feed conversion ratios used to assess nutritional requirements for each type of livestock product. Animal fats derived from livestock in proportion to the energy from meat.	
Demand	Consumption is defined as commodities reaching the consumer, and includes subsequent losses or waste, the rate of which varies globally.	Countries with higher rates of losses and waste at the consumer therefore leads to greater consumption and higher HALF values.
Demand	Diets considered are adopted on average by the global population.	Stylised scenarios of global adoption of selected diets.

Table S3. List of commodities tracked with global mean consumption rates per capita and total agricultural area used in production in 2011, ordered by decreasing energy intake.

Commodity	Commodity type	Commodity	Commodity type
Apples	Primary crop	Olives	Primary crop
Bananas	Primary crop	Onions	Primary crop
Barley	Primary crop	Oranges and mandarines	Primary crop
Beans	Primary crop	Palm Oil	Processed commodity
Beer	Processed commodity	Palmkernel Cake	Processed commodity
Beverages Alcoholic	Processed commodity	Palmkernel Oil	Processed commodity
Beverages Fermented	Processed commodity	Peas	Primary crop
Bovine Meat	Animal product	Pigmeat	Animal product
Cassava	Primary crop	Pimento	Primary crop
Cereals Other	Primary crop	Pineapples	Primary crop
Citrus Other	Primary crop	Plantains	Primary crop
Cocoa Beans	Primary crop	Potatoes	Primary crop
Coconut Oil	Processed commodity	Poultry Meat	Animal product
Coconuts - Incl Copra	Primary crop	Pulses Other	Primary crop
Coffee	Primary crop	Rape and Mustard Cake	Processed commodity
Copra Cake	Processed commodity	Rape and Mustard Oil	Processed commodity
Cotton lint	Processed commodity	Rape and Mustardseed	Primary crop
Cottonseed	Processed commodity	Rice	Primary crop
Cottonseed oil	Processed commodity	Ricebran Oil	Processed commodity
Cottonseed cake	Processed commodity	Roots Other	Primary crop
Dates	Primary crop	Rye	Primary crop
Eggs	Animal product	Seed cotton	Primary crop
Fats animal	Animal product	Sesame seed	Primary crop
Fish Seafood	Primary crop	Sesameseed Cake	Processed commodity
Fruits Other	Primary crop	Sesameseed Oil	Processed commodity
Grapefruit	Primary crop	Sorghum	Primary crop
Grapes	Primary crop	Soyabean Cake	Processed commodity
Groundnut Cake	Processed commodity	Soyabean Oil	Processed commodity
Groundnut Oil	Processed commodity	Soyabeans	Primary crop
Groundnuts	Primary crop	Spices Other	Primary crop
Lemons Limes	Primary crop	Sugar beet	Processed commodity
Maize	Primary crop	Sugar beet	Primary crop
Maize Germ Oil	Processed commodity	Sugar cane	Primary crop
Meat Other	Animal product	Sugar non-centrifugal	Processed commodity
Milk and products	Animal product	Sunflower seed	Primary crop
Millet	Primary crop	Sunflowerseed Cake	Processed commodity
Molasses	Processed commodity	Sunflowerseed Oil	Processed commodity
Mutton & Goat Meat	Animal product	Sweet potatoes	Primary crop
Nuts	Primary crop	Sweeteners Other	Processed commodity
Oats	Primary crop	Tea (including mate)	Primary crop
Oil palm fruit	Primary crop	Tomatoes	Primary crop
Oilcrops Oil Other	Processed commodity	Vegetables Other	Primary crop
Oilcrops Other	Primary crop	Wheat	Primary crop
Oilseed Cakes Other	Processed commodity	Wine	Processed commodity
Olive Oil	Processed commodity	Yams	Primary crop

Table S4. Primary crops to area allocation ratios of commodities produced by processing, for global quantities processed >10Mt in 2011.

Primary Crop	Commodities produced	Area	Crop
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		allocation ratio	processed in 2011 (Mt)
Sugar cane	Sugar : Molasses : Sugar non-centrifugal	70 : 26 : 4	1271.2
Sugar beet	Sugar	100	247.1
Oil palm fruit	Palm Oil : Palmkernel Cake : Palmkernel Oil	78 : 11 : 11	245.6
Soyabeans	Soyabean oil : Soyabean cake	50 : 50	224.1
Seed cotton	Cotton lint: Cottonseed	85 : 15	78.7
Rape and Mustardseed	Rape and Mustard Oil : Rape and Mustard Cake	50 : 50	57.3
Maize	Sweeteners Other : Maize Germ Oil	90 : 10	44.9
Grapes	Wine	100	39.8
Sunflower seed	Sunflowerseed oil : Sunflowerseed cake	50 : 50	31.7
Cottonseed	Cottonseed oil: Cottonseed cake	50 : 50	31.1
Barley	Beer	100	25.9
Coconuts - Incl Copra	Coconut oil : Copra Cake	50 : 50	23.4
Olives	Olives oil	100	16.0
Groundnuts	Groundnut oil : Groundnut cake	50 : 50	14.3

SI Results

Table S5. Human Appropriation of Land for Food (HALF) index by country in 2011.

Country	HALF	Country	HALF	Country	HALF
Mozambique	9.43	Chad	28.19	Lebanon	48.50
Liberia	9.54	China	28.56	Azerbaijan	48.51
Bangladesh	10.60	Trinidad and Tobago	29.04	Lithuania	49.27
Sri Lanka	11.49	Djibouti	29.13	Paraguay	50.00
Togo	12.58	El Salvador	29.23	Bolivia	50.04
Sierra Leone	12.79	Iran	29.52	Latvia	50.20
Malawi	12.88	Grenada	29.80	Bahamas	50.87
Indonesia	12.97	Japan	30.33	Czech Republic	52.38
Iraq	14.82	Georgia	31.35	Central African Republic	52.86
Thailand	14.83	Saint Lucia	31.80	Estonia	52.90
DPR of Korea	15.08	Morocco	32.22	Kyrgyzstan	53.36
Benin	15.71	Dominican Republic	32.27	Oman	53.38
India	15.76	Mauritius	32.37	Costa Rica	53.41
Zambia	15.97	Suriname	32.43	Ecuador	53.56
Cambodia	16.03	Jordan	33.02	Panama	53.75
Guinea-Bissau	16.75	Tunisia	33.25	Chile	54.54
Guinea	17.00	Saudi Arabia	33.27	Montenegro	54.62
Haiti	17.03	Bulgaria	33.28	New Caledonia	55.28
Nigeria	18.26	Burkina Faso	33.67	Spain	56.34
Ghana	18.64	Kenya	34.66	Armenia	56.50
Solomon Islands	18.93	Algeria	34.79	Belarus	57.89
Rwanda	18.99	World	35.13	Russian Federation	58.03
Sao Tome and Principe	19.00	Cuba	35.22	Venezuela	62.45
Ethiopia	19.10	Swaziland	35.62	Albania	64.65
Cote d'Ivoire	19.48	Pakistan	35.65	Uzbekistan	65.77
Philippines	19.84	Ukraine	37.27	United Kingdom	67.21
Gambia	19.88	Samoa	37.40	Portugal	68.21
Yemen	19.95	Bosnia and Herzegovina	37.87	Germany	70.10
Peru	20.25	Egypt	37.92	Malta	70.48
Kiribati	20.69	Namibia	38.25	Netherlands	70.92
Tajikistan	20.75	United Arab Emirates	39.02	Slovenia	71.21
Republic of Moldova	20.75	Serbia	39.21	Belgium	74.16
United Republic of Tanzania	21.05	Macedonia	39.23	Israel	74.22
Angola	21.08	Fiji	39.29	Uruguay	74.26
Madagascar	21.21	Slovakia	39.61	Ireland	75.10
Lao PDR	21.33	Botswana	39.80	Norway	75.82
Guyana	21.38	Cyprus	39.84	Finland	76.93
Guatemala	21.43	Saint Kitts and Nevis	39.90	Kazakhstan	77.04
Congo	21.43	St Vincent and the Grenadines	40.03	Switzerland	79.14
Myanmar	21.82	Poland	40.04	Austria	79.72
Uganda	22.35	Brunei Darussalam	40.33	Greece	81.07
Nicaragua	22.57	Dominica	40.61	Italy	82.82
Malaysia	22.74	Republic of Korea	41.30	Iceland	85.02
Senegal	22.82	Gabon	41.71	French Polynesia	86.51
Zimbabwe	23.44	South Africa	42.79	France	86.72
Afghanistan	23.84	Turkey	43.26	Turkmenistan	88.05
Cameroon	25.15	Niger	43.84	Brazil	88.18
Lesotho	26.09	Hungary	44.06	Sweden	89.71
Viet Nam	26.18	Mali	44.14	Luxembourg	89.82
Nepal	26.63	Colombia	44.77	Canada	92.16
Jamaica	26.71	Antigua and Barbuda	46.07	Denmark	96.60
Belize	26.79	Romania	46.32	United States of America	97.73
Maldives	27.61	Mexico	46.39	Mongolia	99.29
Honduras	27.98	Mauritania	47.00	Bermuda	110.75
Vanuatu	28.04	Kuwait	47.38	Australia	112.20
Cabo Verde	28.06	Barbados	47.58	Argentina	114.88
Timor-Leste	28.13	Croatia	48.25	New Zealand	135.76

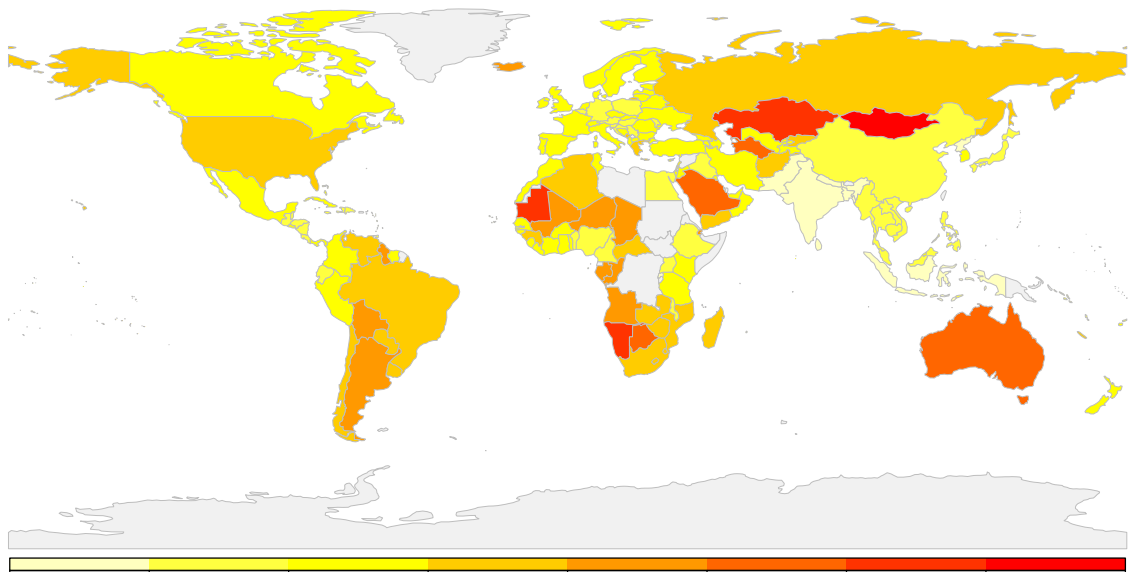
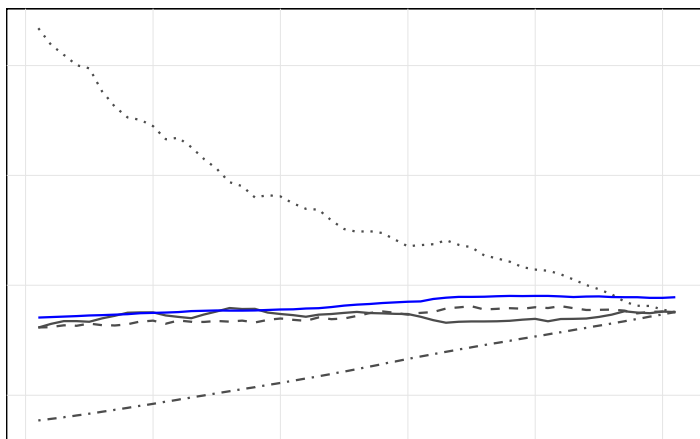


Figure S1. Agricultural land use footprint for food consumption (ha/capita) per country, including land domestic production and displacement through international traded. Following the approach of (Jalava et al. (2014) and Alexander et al. (2015), net exports are accounted for as domestically produced by reducing land allocations on a proportional basis, while net imports are included at the global mean yield of exports, weighed by net export quantities.



al land

Figure S2. Global HALF index from 1961 to 2011 for variations in diet, population and production efficiencies individually and combined, and the total agricultural land area expressed as a fraction of the world's land surface.

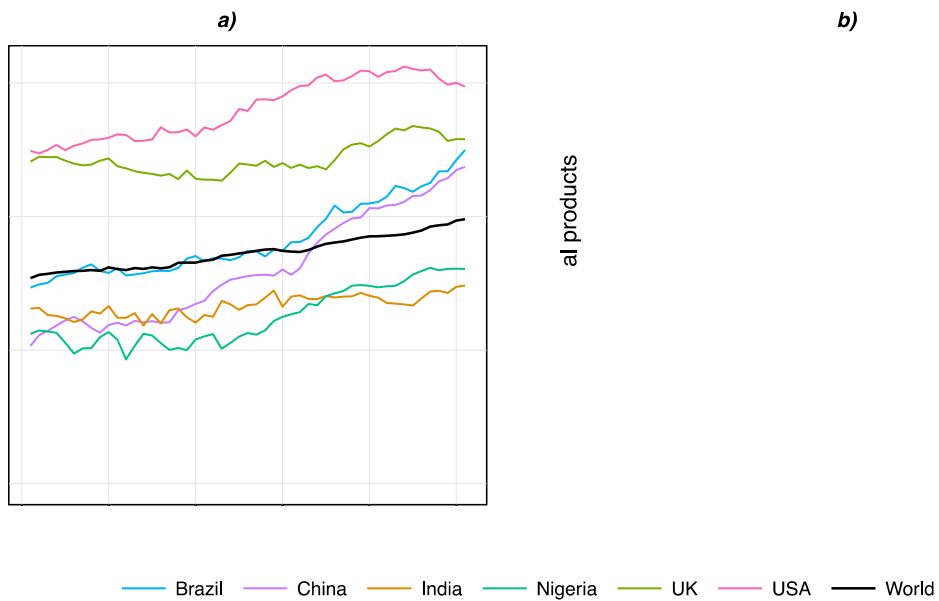


Figure S3. Mean protein per capita and percentage derived from animal productions in foods consumed from 1961 to 2011 globally, and for selected countries. This includes commodity wastage at the consumer, but not in the food supply chain.